

# Winter Bird Use of Urban and Rural Habitats in Ontario

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Smith, Paul G. R. 2003. Winter bird use of urban and rural habitats in Ontario. *Canadian Field-Naturalist* 117(2): 173-183.

This study examined variation in winter bird diversity, abundance, biomass and species and guild composition on plots in farmland, rural forests, mixed rural habitat, cities and three categories of urban natural areas in southern Ontario. Small (<20 g), insectivorous, upper canopy, bark foraging and forest species were more prevalent in rural mixed habitat, urban natural areas and especially rural forests. Larger (20-100g), ground feeding and omnivorous birds typified cities and farms. Many bird species classified as forest, forest edge and field species during summer also favored these same habitats during winter. The 52 sample plots can be accurately classified into the seven habitat categories through discriminant analysis using the percentages of small species, insectivores, forest species and alien species. Urban natural areas, forests and mixed rural habitat showed the highest diversity while cities and farms showed the lowest diversity. Both bird abundance and biomass were highest in urban natural areas with open water and urban residential and commercial areas, whereas both were lowest in farmland.

**Key Words:** bird populations, urban ecology, winter, urban-rural ecological gradient, Ontario.

The landscapes of Ontario south of the Canadian Shield, like much of eastern North America, are mosaics of cities, farms, forests and successional habitats. The effects of urbanization, farming and fragmentation on breeding birds have been widely studied. This is less true of wintering birds (Erskine 1980; 1992; Hohtola 1978; Lancaster and Rees 1979; Blake 1987; Tilghman 1987; Telleria and Santos 1995; Morneau et al. 1996). Understanding the effects of landscape change on birds depends on knowledge of effects of habitat on birds during all seasons. This paper compares and contrasts winter bird use of seven urban and rural habitat categories in terms of avian diversity, abundance, biomass, bird size, guilds and species composition.

Urban ecosystems generally have lower biological diversity compared to forests and rural habitats (Gilbert 1989; Adams 1994; Marzluff et al. 1998). However, along with the lower diversity, urban bird assemblages often show greater abundance and biomass than do those in forests (Erskine 1980; Hohtola 1978; Lancaster and Rees 1979; Adams 1994).

Farmland is a mosaic of habitats for birds, but many farmland birds are edge species associated with hedgerows and remnant natural vegetation, and other typically urban species may be associated with farm buildings (Rodenhouse et al. 1995). Relatively few species are specifically associated with croplands (Møller 1984). Nevertheless, some croplands provide considerable food in soil invertebrates and seed (Lack 1992). Areas with more diversity in farm habitats and in farming practices show far greater diversity of bird life (Parrish et al. 1994; Bignal and McCracken 1996; Newton 1998).

The effect on bird life of habitat fragmentation from farming and urbanization is now a central issue in conservation biology (Whitcomb et al. 1981; Ambuel

and Temple 1983; Robbins et al. 1989; Saunders et al. 1991; Hinsley et al. 1996; Nour et al. 1999). The affinities of breeding species for field, forest edge and forest habitats are now well established. A few studies suggest that similar affinities exist for wintering bird species (Blake 1987; Tilghman 1987; Telleria and Santos 1995).

Winter and breeding bird assemblages differ in many respects. In temperate regions, bird species-richness and abundance are lower during winter, due to the absence of neotropical and short-distance migrants (Morrison et al. 1987). Snow cover and depth limit ground foraging (Erskine 1980). Winter is a critical time for many resident species, and winter events can affect population levels during the breeding season (e.g., Mehlman 1997). Patterns of species presence have changed not only due to habitat modifications, but also due to increases in bird feeders in urban and rural areas (Root 1988; Wilson 1994).

Periodic "eruptions" of northern seed-eating and raptorial species in particular years during winter can dramatically increase local densities of birds and change species composition (Bock and Lepthien 1976; Root 1988; Hochachka et al. 1999). Intra-species and inter-species flocking is important in winter bird assemblages. This phenomenon relates to the reduced sedentary and territorial behavior during winter and is thought to have adaptive significance (e.g., Berner and Grubb 1985; Morrison et al. 1987).

While many aspects of breeding bird use of different urban and rural habitat categories have been thoroughly studied, research on winter birds is sketchy. Therefore, this research addressed the following specific questions. Are the well-documented differences in diversity, abundance, biomass and species composition among rural and urban breeding bird communities similar for

wintering birds in southern Ontario? How does the more limited range of food types during winter affect the guilds and sizes of birds using different habitat categories? Are the now familiar preferences of breeding bird species for forest, forest edge and field evident in winter habitat use? To address these questions, this study analyzed differences in winter bird use of different habitat categories in rural and urban areas in southern Ontario, Canada. It examined variation in winter bird diversity, abundance, biomass and species and guild composition on plots in farmland, forests, mixed rural habitat, cities and three-categories of urban natural areas.

## Methods

### Study Plots

The Winter Bird Population Study (WBPS) is a method for sampling the winter birds of defined census plots (Kolb 1965; Robbins 1972, 1981). Eight to ten, and occasionally fewer, counts on each fixed-area plot are conducted in December through February. The WBPS method can be robust to the use of different observers on different plots (Smith 1984a).

Fifty-two plots from across southern Ontario were used in this study (Appendix), drawn from those published in *American Birds* and *Ontario Field Biologist*. Further description of some plots can be found elsewhere (Campbell and Dagg 1976; Erskine 1975; Smith et al. 1981, 1982; Smith 1984a, b). Climate varies across the geographic range of locations of these study plots and influences the distribution and abundance of a number of species (Erskine 1980). Climate and other factors contribute to the large year-to-year variation in the winter occurrence and abundance of many species (Bock and Lepthien 1976; Bock and Root 1981; Root 1988; Smith 1984a).

### Habitat Classification

The study plots reflect different habitat categories in urban and rural southern Ontario and were classified into seven categories—three rural and urban categories (cf., Brady et al. 1979). These reflect gradients of human modification through urbanization and agriculture. Table 1 summarizes the classification of habitat categories. This classification has significant limitations but provides a basis for assessing some broad scale patterns.

Fourteen of the plots are drawn from a range of rural landscapes including rural forest, mixed rural habitats and farmland.

Seven rural forested plots included deciduous, mixed and coniferous forests (1-7, in Appendix). Six of the plots were within larger continuous protected habitats including conservation lands, Rondeau Provincial Park and St. Lawrence Islands National Park.

Four plots were composed of mixed rural habitats including forest, old field (successional), and agricultural field (8-11, in Appendix).

Three plots composed of farmland were included (50-52 in the Appendix).

Thirty-three plots reflect a range of urban habitat categories including urban and suburban residential, commercial and urban natural areas including parks, urban open space, woodlands and other natural or semi-natural areas within cities and towns. These plots reflect different levels of disturbance along an urbanization gradient.

Twenty-seven plots of urban natural areas included ravines, valleys, woodlots, and parks. Urban natural areas were divided into three categories as described in Table 1. Five plots in urban natural areas with more than 50 m<sup>2</sup> open water during the winter were classified as a separate category, urban natural with water (12-16 in Appendix). The presence of open water is known to significantly influence abundance, diversity and species composition of winter bird assemblages (Tilghman 1987; Gilbert 1989; Adams 1994). The other 22 plots in urban natural areas were divided into those more or less than 10 ha in area (Table 1). Twelve plots were > 10 ha (17-28 in Appendix), while a further ten were < 10 ha (29-38 in Appendix). This division reflects the fact that area is known to significantly influence the diversity and species composition of winter bird assemblages (Blake 1987; Tilghman 1987; Telleria and Santos 1995). Exploratory data analysis of these 27 plots indicated that area and presence of open water were important variables in explaining variance among the 27 plots and therefore a useful means to categorize the urban natural areas.

Urban areas or "cities" included two downtown plots, four plots in older, high-density residential parts of three cities, and one plot in a small town (39-49 in the Appendix). Four other plots were located in suburban, lower-density portions of two cities.

### Variables Selected for Analysis

A variety of measures of avian community structure and composition were used to examine variation across the habitat categories. Winter bird diversity, abundance, biomass, mass, guild structure and species composition were all examined.

Ecological and avian diversity have been measured in a wide variety of ways to address different definitions of diversity and methodological issues (Magurran 1988). Consequently, several indices of diversity were employed in this study. The number of species recorded was the measure of species richness. Species richness is known to have inherent and sampling-based biases (Hill 1973; James and Rathbun 1982; Magurran 1988) and so other indices of diversity were used. Simpson's diversity index,  $N_2$  ( $N_2 = 1/\sum p_i$ , where  $p_i$  is the proportion of each species), a common measure of diversity (Hill 1973), was also used. Plot area was used as a covariate in analyzing species richness and Simpson's diversity index [ $N_2$ ] to correct for the known correlation of diversity with plot area (James and Rathbun 1982; Magurran 1988).

TABLE 1. A description of the habitat categories used in this study.

Urban / Rural	Habitat Category	Description
Rural Areas	Forest	Naturally occurring deciduous, coniferous or mixed forest in a rural landscape consisting primarily of natural habitats and farmland.
Rural Areas	Mixed Rural	Plots that include more than one of farmland, successional field, wetland and forest in a rural landscape consisting primarily of natural habitats and farmland.
Rural Areas	Farm	Lands used for crops or pasture with little significant natural vegetation except hedgerows, occurring in a rural landscape consisting of farmland and natural habitats. No farm buildings were present on these plots.
Urban Areas	Urban Natural	Public open spaces with a mixture of natural habitats and horticultural vegetation, surrounded by residential, commercial or industrial areas and located within cities with > 50,000 people. Sites in this category were > 10 ha in extent. The WPBS plot generally occupied the full extent of the natural area or park.
Urban Areas	Urban Natural with Open Water	As for Urban Natural, with open water at least 50 m <sup>2</sup> in extent on a creek, river or pond.
Urban Areas	Small Urban Natural	As for Urban Natural, but ≤10 ha in extent.
Urban Areas	City	Commercial or residential areas within cities or small towns and dominated by buildings and pavement. The sites vary in the density of buildings, the percentage of area covered by pavement and the amount of horticultural vegetation present. Both higher-density urban core areas and lower-density suburban areas are included.

Rarefaction is a technique for comparing collections of items with unbalanced sample sizes, which has always been a problem in measuring ecological diversity (Engstrom and James 1981; James and Rathbun 1982; Brewer and Williamson 1994; Gotelli and Colwell 2001). In this study, the expected number of species in a sample of five birds [ $E(S_5)$ ] was used as the rarefaction index of diversity (James and Rathbun 1982).

Total bird abundance, total land bird abundance and the abundance of each species were assessed as the number of birds per hectare, consistent with standard methods for the Winter Bird Population Study (Kolb 1965; Robbins 1972, 1981). For total abundance, the number of survey hours per hectare was used as a covariate to correct for sampling effort (similar to the suggestion of Bock and Root 1981).

Biomass and average bird size or mass for each plot were calculated in a manner similar to Mauer (1985) using the average mass values for bird species compiled by Dunning (1993). The average mass for each species was multiplied by the average number of birds of that species found on the plot. Total biomass was summed across all species and divided by the plot area in hectares. Similarly, average bird mass was calculated as the total biomass divided by the total number of birds.

The occurrence of birds in different size classes was examined by determining for each plot the percent-

age of birds in four size classes: < 20 g; 20-40 g; 40-100 g; and > 100 g.

The percentages of birds in different feeding and substrate guilds were estimated using the guild classifications of DeGraaf et al. (1985). Note that a species may be classified in more than one guild. The percentage of field, forest edge and forest species was calculated based on the species' affinities noted by Cadman et al. (1987), Whitcomb et al. (1981), Root (1988), and Freemark and Collins (1992). These are classifications based on breeding season preferences. In this paper I test the applicability of such classifications to winter bird habitat analyses.

#### Statistical Methods

Analysis of variance was used to test for significant differences in bird community variables among habitat categories. For sites with more than one year of sampling, the mean values for each plot were used to avoid the problem of pseudoreplication (Hulbert 1984). Where departures from the normal distribution and homogeneity of variances could not be overcome through transformation, the non-parametric Kruskal-Wallis analysis of variance test was applied.

## Results

### Diversity

Bird species richness, Simpson's diversity index ( $N_2$ ) and the rarefaction diversity index [ $E(S_5)$ ] all differed significantly ( $P < 0.0001$ ,  $P < 0.01$  and  $P < 0.001$

respectively) among the habitat categories (Table 2). Urban natural areas, forests and mixed habitats showed the highest avian diversity while city and farm showed the lowest diversity (Table 2). Larger urban natural areas (> 10 ha) had diversity indices 5-19% higher than forests.

City areas supported half as many species and about 70% the levels of Simpson's diversity index,  $N_2$ , as forests. The rarefaction diversity index  $[E(S_2)]$  in city areas was about 75% of that for forests. Farmland diversity levels were about one-third those for forests, except the rarefaction diversity index  $[E(S_2)]$  which was 69% the value for forests. Species richness in mixed rural habitats was about 20% lower than in rural forests, whereas other indices were about as high or higher than values in forests (Table 2).

#### Abundance, Biomass and Size

Total bird abundance (Table 2) was significantly different ( $P < 0.01$ ) among the habitat categories, due to particularly high abundances in two habitat categories. Total bird abundance was highest in urban natural areas with open water and residential and commercial city areas. Farmland and rural mixed habitats had the lowest levels. Forests and urban natural areas had similar, intermediate levels of bird abundance.

Total bird biomass was significantly different among habitat categories ( $P < 0.001$ ). The highest values were recorded in urban natural areas with open water, due to the presence of waterfowl, and in city areas, due to higher densities and larger species such as the Rock Dove (*Columba livia*). Other urban natural areas and forest had intermediate biomass levels (Table 2).

In urban natural areas with open water, total abundance was 80% higher than in rural forests. Total biomass was 13 times greater in urban natural areas with open water than in rural forests (Table 2). Other urban natural areas showed bird abundance and biomass levels similar to rural forests. Both farms and mixed rural habitats showed abundance levels roughly one third of rural forests. Biomass on farmland was about 30% that in forests while biomass in mixed rural habitats was about 60% of that in forests. Mean bird size was roughly 40% higher in farmland compared with forests. Total bird abundance in cities was 46% higher than that recorded in forests. Biomass was 90-100% higher in cities than in forests.

Significant differences ( $P < 0.001$ ) among habitat categories in the occurrence of birds of different sizes (Figure 1) indicate the cause of the trends of differences in overall biomass shown in Table 2. A preponderance of small bird species (< 20 g) is indicative of forested habitats, as is a relatively even mix of the other size classes. Mixed rural habitats differed from forests primarily in having additional birds in the >100 g class, specifically more raptors. In fact, rural mixed habitats had the highest average bird size. In contrast, city areas showed a preponderance of birds in the intermediate 20-40 gram and 40-100 gram ranges.

TABLE 2. Bird species richness, rarefaction diversity index  $[E(S_2)]$ , Simpson's diversity index  $[N_2]$ , abundance and biomass levels in seven categories of urban and rural habitats.

Variable	Rural Areas					Urban Areas			Statistical Significance <sup>1</sup>
	Farm	Mixed Rural	Forest	Urban Natural	Small Urban Natural	Urban Natural with Open Water	City		
Species richness	Mean 4.5 Range 4-7	Mean 12.3 Range 9-15	Mean 14.7 Range 10-19	Mean 15.7 Range 10-20	Mean 7.8 Range 1-16	Mean 19.2 Range 10-25	Mean 6.8 Range 3-15		****2
Rarefaction index of diversity $[E(S_2)]$	Mean 2.42 Range 2.0-3.1	Mean 3.5 Range 3.5-3.5	Mean 3.52 Range 3.1-4.0	Mean 3.69 Range 2.4-4.2	Mean 3.55 Range 2.4-4.1	Mean 2.88 Range 1.8-4.3	Mean 2.66 Range 2.4-3.1		***
Simpson's index of diversity $(N_2)$	Mean 2.3 Range 1.3-2.5	Mean 5.6 Range 3.6-9.4	Mean 5 Range 4.3-6.9	Mean 6.1 Range 2.2-9.2	Mean 3.3 Range 1.0-6.5	Mean 5 Range 1.5-12.5	Mean 3.3 Range 2.0-3.6		**2
Total abundance (birds / ha)	Mean 0.72 Range 0.11-1.41	Mean 0.765 Range 0.25-1.48	Mean 2.39 Range 1.82-3.57	Mean 2.46 Range 0.75-4.99	Mean 2.45 Range 0.31-16.07	Mean 4.4 Range 1.88-4.37	Mean 3.5 Range 0.04-6.67		**2
Biomass (g / ha)	Mean 37.5 Range 4.5-235.1	Mean 92.9 Range 23.5-333.9	Mean 142.4 Range 64.0-288.2	Mean 164.9 Range 52.9-390.5	Mean 111.1 Range 13.9-677.1	Mean 1895 Range 246.5-13,210.9	Mean 271.3 Range 5.4-899.5		***?

<sup>1</sup> Statistical significance from Analysis of Variance F-test: \* =  $P < .05$ ; \*\* =  $P < .01$ ; \*\*\* =  $P < .001$ ; \*\*\*\* =  $P < .0001$ .

<sup>2</sup> Plot area was used as a covariate for species richness and  $N_2$ , and sampling effort (hours / ha) as a covariate for abundance and biomass.

The latter category includes the House Sparrow (*Passer domesticus*) and Starling (*Sturnus vulgaris*). Farm areas had a predominance of birds in the 40-100 gram range, which includes the Snow Bunting (*Plectrophenax nivalis*) and Eastern Meadowlark (*Sturnella magna*) (Figure 1).

#### Guilds

Of food-based feeding guilds, the percentage of insectivorous birds differed most (statistical significance  $P < 0.0001$ ) among habitats (Table 3). Forests, with 25.3 percent, had the highest percentage of insectivores, followed by farms (19.1 percent) and mixed rural habitats (16.7 percent). Cities had only 1.5 percent insectivorous birds. This pattern shows small insectivorous bird species were more indicative of natural, rural forest conditions. Rural mixed habitat shows the highest proportion of carnivorous birds. No significant differences were found among habitat categories for the percentages of omnivores, granivores, herbivores or frugivores (Table 3).

The percentage of four substrate-based guilds differed significantly ( $P < 0.01$ ) among habitat categories (Table 3). Ground-gleaning and foraging species predominated in all urban and farm habitats. Lower canopy species were also important in urban natural areas and rural mixed habitats. In forests, a more even distribution is evident with significant proportions of birds that forage in the upper and lower canopy, on tree trunks and on the ground.

#### Species Composition

Species showing significant differences ( $P < 0.05$ ) in abundance among the seven habitat categories are shown in Table 4. Of the fifteen most common species only Pine Siskin (*Carduelis pinus*), Evening Grosbeak (*Coccothraustes vespertinus*), and Cedar Waxwing (*Bombycilla cedrorum*) failed to show significant differences among habitat categories, all species subject to periodic extreme peaks in winter abundance (Bock and Lepthien 1976; Root 1988).

Predictably the three alien species — House Sparrow, European Starling and Rock Dove—occurred almost exclusively in city and urban natural areas (Table 4). Some other species reached their greatest abundances in urban natural areas (Table 4), including Mourning Dove (*Zenaidura macroura*), American Robin (*Turdus migratorius*), and Northern Cardinal (*Cardinalis cardinalis*).

Species associated with forest edge and fields are the most common in this study. The species American Robin through White-breasted Nuthatch (*Sitta carolinensis*) in Table 4 are generally regarded as either field or forest-edge species. The proportion of forest-edge species is fairly constant at roughly half of birds in forest, mixed rural and urban natural area habitats.

Brown Creeper (*Certhia americana*), Golden-crowned Kinglet (*Regulus satrapa*), Red-breasted Nuthatch (*Sitta canadensis*), Ruffed Grouse (*Bonasa umbellus*), and Hairy Woodpecker (*Picoides villosus*) showed the clearest affinities for forest habitats (Table 4). Fourteen other species generally regarded as forest species during the breeding season occurred at frequencies too low to reveal statistical trends. One third of winter birds recorded in forest areas are regarded as forest species, a much higher percentage than for all other habitats ( $P < 0.001$ ). This analysis suggests that a suite of bird species can be identified that preferentially use forest during winter, as well as during the breeding season.

#### Differences Among Habitat Categories

A series of significant ecological gradients is revealed in the percentages of small birds, insectivorous species, forest species and alien species across the seven habitat categories in urban and rural southern Ontario. These four variables can be used to summarize the differences in avifauna of the seven habitat categories using multivariate analysis (Table 5). The percentages of birds  $< 20$  g in mass, of insectivorous species, of alien species and of forest species separate

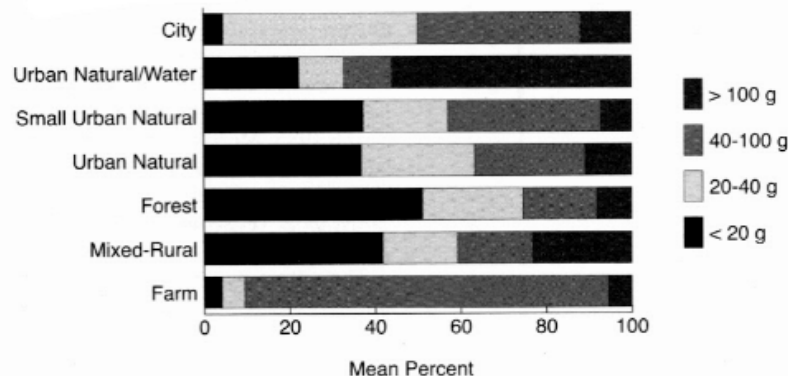


FIGURE 1. The distribution of birds in four size classes in seven categories of urban and rural habitats. Analysis of variance showed highly significant differences ( $P \leq 0.001$ ).

TABLE 3. Percentages of birds in different guilds in seven categories of urban and rural habitats.<sup>1</sup>

Guild Type	Rural Areas				Urban Areas				Statistical Significance <sup>2</sup>
	Guild	Farm	Mixed-Rural	Forest	Urban Natural	Small Urban Natural	Urban Natural with Open Water	City	
Food	Insectivore	19.1	16.7	25.3	13.9	10.8	10.3	1.5	****
	Carnivore	2.7	10.1	1.4	2.3	3.3	2.5	0.0	*
	Omnivore	20.4	39.4	38.3	43.8	53.1	52.2	41.1	
	Granivore	50.1	33.2	35.5	42.7	35.8	33.6	56.4	
	Frugivore	6.0	9.8	15.3	8.2	7.2	8.1	2.5	
	Herbivore	3.5	3.4	3.5	0.7	0.4	2.6	0	
Substrate	Upper Canopy	0.0	16.6	34.9	12.4	7.0	7.6	13.8	***
	Lower Canopy	10.3	43.9	46.2	35.2	31.0	30.7	3.6	**
	Ground	72.6	46.9	35.3	63.7	68.4	68.0	94.6	****
	Bark	19.1	15.8	23.6	13.4	10.7	10.5	0.9	***

<sup>1</sup> Note that numbers for each guild type in each column do not sum to 100% as some species may be members of more than one guild (see DeGraaf et al. 1985).

<sup>2</sup> Statistical significance from Kruskal-Wallis Test is: \* =  $P < .05$ ; \*\* =  $P < .01$ ; \*\*\* =  $P < .001$ ; \*\*\*\* =  $P < .0001$ .

the seven habitat categories in a non-parametric, Epanechnikov kernel discriminant analysis (Hand 1982). The technique successfully categorized 96.2 percent of the 52 plots ( $P < 0.0001$ ). Only two plots were mis-categorized and these were two urban natural area plots classified in one of the other two categories of urban natural areas.

### Discussion

Patterns revealed here regarding winter bird species richness and diversity in urban and rural habitats were similar to other studies in different climatic zones (Lancaster and Rees 1979; DeGraaf and Wentworth 1981, 1986; Tilghman 1987; Adams 1994). Diversity was higher in forest, rural mixed habitats and urban natural areas compared with cities and farms. Tilghman (1987) found species richness in an urban area in Massachusetts, U.S. was only 50% of that in forests and a similar finding is reported here, in a colder region with more consistent snow cover.

In northern and temperate ecosystems, bird species richness and abundance are lower in winter than summer (Anderson et al. 1981; Morrison et al. 1987). The magnitude of the seasonal difference in avian diversity and abundance also differs among habitats. Some studies show that in winter, birds favor areas with greater vegetation cover compared to summer (Morrison et al. 1987, western U.S.) and areas with milder microclimates (Shields and Grubb 1976, north central U.S.). Urban areas provide larger, more stable food supplies, higher temperatures and reduced temperature variability which may result in higher avian productivity and survival rates (Marzluff et al. 1998).

Total bird abundance and biomass were highly variable and differences among habitat categories did not explain much of the variance. The presence of water and riparian bird species had a major influence on total avian abundance in this study. Waterfowl are known to move into cities in the winter due to the higher temperatures and resulting open water (Tilghman 1987; Gilbert 1989; Adams 1994). The occasional occurrence of flocks of northern finches has been noted as a possible cause of low explanatory power of habitat variables related to bird abundance (Tilghman 1987). Intra-species and inter-species flocking, known to be important in winter bird assemblages (e.g., Berner and Grubb 1985; Morrison et al. 1987), might also contribute to greater variability. Abundance is also affected by landscape effects at larger scales than individual study plots (Pearson 1993; Hostletler and Holling 2000).

The differences in the sizes of birds present in different habitat categories are remarkable and this has not been examined to any extent in North America. Hostletler and Holling (2000) found that birds of different sizes responded to the amount of urban forest cover at different scales. In this study, smaller birds (< 20 g) favored forests, mixed rural habitats and

TABLE 4. Densities of individual species in seven categories of urban and rural habitats. Values shown are birds per hectare.

Species	Species' Habitat Affinities <sup>2</sup>	Rural Areas			Urban Areas				City	Significance <sup>1</sup>
		Farm	Mixed-Rural	Forest	Urban Natural	Small Urban Natural	Urban Natural with Open Water	Urban Natural		
House Sparrow <i>Passer domesticus</i>	Urban				0.206	0.462	0.084	1.503	****	
European Starling <i>Sturnus vulgaris</i>	Urban		0.01		0.387	0.638	0.296	0.819	***	
Rock Dove <i>Columba livia</i>	Urban		0.013		0.001		0.044	0.307	***	
American Robin <i>Turdus migratorius</i>	Field				0.021	0.001	0.069	0	*	
Mourning Dove <i>Zenaidura macroura</i>	Field		0		0.145	0.104	0.050		**	
American Tree Sparrow <i>Spizella arborea</i>	Field		0.079	0.044	0.105	0.048	0.014		*	
American Goldfinch <i>Carduelis tristis</i>	Field		0.091	0.23	0.273	0.523	0.072	0.025	**	
Song Sparrow <i>Melospiza melodia</i>	Forest Edge		0.01	0.006	0.058	0.001	0.033		**	
Red-tailed Hawk <i>Buteo jamaicensis</i>	Forest Edge	0.004	0.024	0.004	0.019	0.020	0.010		***	
Black-capped Chickadee <i>Poecete atricapillus</i>	Forest Edge	0.005	0.191	0.610	0.406	0.851	0.269	0.086	**	
Downy Woodpecker <i>Picoides pubescens</i>	Forest Edge	0.006	0.053	0.154	0.144	0.271	0.087	0.009	***	
Dark-eyed Junco <i>Junco hyemalis</i>	Forest Edge			0.079	0.205	0.276	0.087	0.006	**	
Northern Cardinal <i>Cardinalis cardinalis</i>	Forest Edge		0.01	0.019	0.164	0.243	0.070	0.028	***	
Blue Jay <i>Cyanocitta cristata</i>	Forest Edge		0.025	0.175	0.103	0.115	0.036	0.011	***	
White-breasted Nuthatch <i>Sitta carolinensis</i>	Forest Edge		0.026	0.133	0.074	0.114	0.046	0.029	*	
Hairy Woodpecker <i>Picoides villosus</i>	Forest	0.007	0.025	0.095	0.018	0.068	0.024	0.002	*	
Brown Creeper <i>Certhia americana</i>	Forest		0.038	0.081	0.016	0.004	0.016	0.001	**	
Golden-crowned Kinglet <i>Regulus satrapa</i>	Forest		0.025	0.044	0.007	0.002			***	
Red-breasted Nuthatch <i>Sitta canadensis</i>	Forest			0.125	0.007	0.010			*	
Ruffed Grouse <i>Bonasa umbellus</i>	Forest		0.051	0.076		0.001	0.002		***	

<sup>1</sup> Statistical significance from Kruskal-Wallis Test is: \* = P < .05; \*\* = P < .01; \*\*\* = P < .001; \*\*\*\* = P < .0001.

<sup>2</sup> Classification based primarily on Cadman et al. 1987.

TABLE 5. Percentages of alien, forest, insectivorous and small bird species in seven categories of urban and rural habitats. A non-parametric, Epanechnikov kernel discriminant analysis correctly categorized 96.2 percent of the 52 plots as to habitat classification using these four variables ( $P < 0.0001$ ).

Variable	Rural Areas			Urban Areas			
	Farm	Mixed-Rural	Forest	Urban Natural	Small Urban Natural	Urban Natural with Open Water	City
Percentage Alien Species	0.0	1.9	0.0	21.2	22.1	27.4	80.4
Percentage Forest Species	19.1	21.7	37.9	19.9	14.5	13.9	1.5
Percentage Insectivores	18.0	17.7	33.3	12.3	9.4	8.4	1.2
Percentage Small Bird Species (< 20 g)	4.2	41.7	51.1	36.8	37.3	22.3	4.6

urban natural areas. Larger birds, 20-100 g in size, were prevalent in cities and on farms. Marzluff et al. (1998) speculated that urbanization likely favours ground foraging bird species and discourages canopy and bark foragers. Ground-foraging birds characteristic of open country are generally larger than species foraging on branches in woodland habitats (e.g., Polo and Carrascal 1999). More complex forested habitats are thought to provide greater opportunities for small birds foraging among branches and milder microclimates, especially protection from wind. Woodlands may support a larger variety of sizes of wintering bird species as observed in this study.

Both cities and farms favor a small number of granivorous and omnivorous species (Møller 1984; O'Connor and Shrubbs 1986; Lack 1992; Adams 1994; Marzluff et al. 1998). The volume of food available to birds in some cities is enormous, perhaps more than the entire bird assemblage would require (Lancaster and Rees 1979). Raptor populations are sometimes lower in cities (Tomialojc 1982; Adams 1994) and this study suggests this may be the case in southern Ontario. In this study, granivorous and omnivorous birds were ubiquitous in all habitats, consistent with the available winter food. Significant variation in the proportion of insectivores across the habitat categories was a major trend noted in this study and one variable that best distinguishes the seven habitat categories.

The identification of species' preferences during summer for field, forest edge and forest has been widespread (Whitcomb et al. 1981; Ambuel and Temple 1983; Robbins et al. 1989; Freemark and Collins 1992). But investigation of species' preferences during winter has been more limited (Blake 1987; Tilghman 1987; Telleria and Santos 1995). This study corroborates preliminary work in the northern United States on species' preferences during winter (Blake 1987, Illinois; Tilghman 1987, Massachusetts) and suggests a broader trend across several climatic zones. Many of the species associated with forest, forest edge and field during summer show similar preferences during winter.

Some interesting new findings arise from analysis of winter bird use of urban and rural habitats in southern Ontario, as well as corroboration and extension of results from a few other studies of winter bird use in the northern U.S. (Blake 1987; Tilghman 1987). Small, insectivorous and forest species were prevalent in forests while larger, ground feeding and omnivorous birds overwhelmingly dominated cities and farms. Urban natural areas occupy an intermediate point on this gradient from natural rural forests to most human-modified sites on farms and in cities. Granivorous and omnivorous species dominated all habitat categories. Variation in the abundance of small, insectivorous, upper and lower canopy and forest bird species was one of the major sources of variance in this study. Trends in biomass and bird sizes suggest further research on these aspects of winter bird communities.



The different habitat categories can be successfully classified based on the percentages of small species, insectivores, forest species and alien species using multivariate discriminant analysis. More research on winter bird use of farmland habitats in North America would be useful given the patterns found in European work. Additional work measuring species preferences in winter for forest, forest-edge and field habitats would also be useful given the interesting results so far.

### Acknowledgments

Thanks are extended to the volunteers who conducted the winter bird population studies used in this paper. The Toronto Bird Observatory, James L. Baillie Fund and Toronto Field Naturalists, supported many of these field studies. John Theberge provided support and guidance at early stages of this research. Mike Cadman, Paul Eagles, Anthony Erskine, George Francis, John Theberge and anonymous reviewers provided comments on drafts of the manuscript.

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Received 14 December 2000

Accepted 29 October 2003

APPENDIX. The 52 Winter Bird Population Study plots in southern Ontario used in this paper.

Site #	Site Name	Region County/	Area (ha)	Site #	Site Name	Region County/	Area (ha)
1	Chats Falls	Ottawa-Carleton	8.4	25	Nordheimer Ravine	Toronto	10.1
2	Purpleville A	York	17.4	26	Wigmore Ravine	Toronto	34.0
3	Chesney Bog	Oxford	15.2	27	Sherwood Park	Toronto	18.5
4	St. Lawrence Island National Park	Leeds	12.5	28	Waterloo Park A	Waterloo	24.3
5	Purpleville B	York	10.5	29	Blockline/ Strasbourg	Waterloo	1.2
6	Rondeau Provincial Park	Kent	6.1	30	Campus Woodlot	Waterloo	0.8
7	South Walsingham	Haldimand-Norfolk	10.1	31	Salt Spring Road	Waterloo	0.7
8	Flamboro area	Hamilton-Wentworth	41.7	32	Upper Gerrard Ravine	Toronto	5.0
9	Homer Watson	Waterloo	10.1	33	Old Country Road	Waterloo	0.7
10	Dundas area	Hamilton-Wentworth	6.3	34	Village Woods	Waterloo	0.7
11	Acton area	Halton	4.9	35	Hidden Valley	Waterloo	1.2
12	Victoria Park	Waterloo	22.9	36	Chatsworth Ravine	Toronto	4.4
13	Waterloo Park B	Waterloo	24.3	37	Mabel Davis	York	6.5
14	Bayview Woods	Toronto	20.0	38	Rockway Gardens	Waterloo	2.3
15	Cootes Paradise	Hamilton-Wentworth	13.0	39	Downtown Waterloo	Waterloo	25.0
16	Riverside Park	Waterloo	81.0	40	Downtown Ottawa	Ottawa-Carleton	41.0
17	Steckle Woods	Waterloo	26.2	41	Westboro/ Highland	Ottawa-Carleton	111.0
18	Moore Park	Toronto	13.2	42	Kent/Percy	Ottawa-Carleton	50.6
19	Strathgowan Woods	Toronto	8.1	43	Hemlock/ Hickory	Waterloo	25.0
20	Cedarvale Ravine	Toronto	20.0	44	Guelph	Wellington	4.0
21	Rosedale Valley	Toronto	19.4	45	Carleton Place	Ottawa-Carleton	103.0
22	Park Drive	Toronto	19.6	46	Alta Vista	Ottawa-Carleton	97.1
23	Lakeside Park	Waterloo	14.2	47	Copeland/Belair	Ottawa-Carleton	111.0
24	Humber River - Land	Toronto	10.9	48	Rockcliffe Park	Ottawa-Carleton	87.0
				49	Beechwood	Waterloo	25.0
				50	Vaughan A	York	20.2
				51	Carp River	Ottawa-Carleton	70.8
				52	Vaughan B	York	40.5